

Effects of Grazing on Songbirds, Sage-grouse, and Invertebrates in Central Montana

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TABLE OF CONTENTS

CONTRIBUTORS	2
EXECUTIVE SUMMARY.....	4
INTRODUCTION.....	5
STUDY AREA.....	9
THE SAGE GROUSE INITIATIVE (SGI) GRAZING PROGRAM.....	10
SONGBIRD COMMUNITY AND REPRODUCTION.....	11
GREATER SAGE-GROUSE DEMOGRAPHICS AND HABITAT.....	13
INVERTEBRATE BIOMASS PREDICTIVE SPATIAL LAYER.....	19
LITERATURE CITED.....	24

EXECUTIVE SUMMARY

The sagebrush steppe of the western US is one of the most imperiled ecosystems in North America. Corresponding declines in sagebrush avifauna mirror their disappearing sagebrush habitat. Most notably, the greater sage-grouse (*Centrocercus urophasianus*; hereafter 'sage-grouse') has been extirpated from approximately half of its range since European settlement. Similarly, grassland and shrubland songbirds inhabiting sagebrush habitat are exhibiting precipitous declines, more so than any other avian guild in North America. Since livestock grazing is the dominant land management practice in sagebrush ecosystems, livestock grazing may offer a promising strategy for conserving and improving avian habitat in sagebrush rangelands. Rest-rotation grazing, in particular, may be useful for promoting habitat diversity. However, grazing effects, regardless of grazing regime, vary depending on ecological context. While rest-rotation grazing has been implemented in other ecosystems, its impacts on sagebrush habitat in central Montana are unknown. A decade of data collection (2011-2020) has allowed us to evaluate the long-term and short-term effects of rest-rotation grazing implemented through the Sage Grouse Initiative (SGI) on songbird reproduction and community metrics, sage-grouse vital rates and habitat, and arthropod biomass.

Songbird species richness, composition, diversity, and reproduction metrics in SGI pastures were largely unchanged relative to non-SGI grazing pastures. However, estimates of abundance for the five most common species suggest species-specific responses to the two grazing systems. For instance, thick-billed longspurs (*Rhynchophanes mccownii*) were most abundant on lands using SGI grazing, while observations of western meadowlarks (*Sturnella neglecta*) were higher on non-SGI plots during the early years of our study. For our three focal species, Brewer's sparrow (*Spizella breweri*), vesper sparrow (*Pooecetes gramineus*), and thick-billed longspur, we located 40% of nests on lands using non-SGI grazing, compared to 60% of nests on lands using SGI grazing. Estimated nest density was higher on SGI grazing plots for thick-billed longspur while nest densities for both Brewer's and vesper sparrows were higher on non-SGI plots. For all focal species, nest success showed little difference between SGI and non-SGI. Both nest density and nest success varied annually. Future work will investigate environmental factors influencing sagebrush steppe songbird reproduction through the development, testing, and implementation of a novel songbird nest success and abundance model.

Findings related to sage-grouse demographics and habitat are preliminary, but thus far indicate minor differences between SGI rest-rotation grazing and non-SGI grazing regimes. SGI plots tended to have greater herbaceous vegetation than non-SGI plots. However, this difference was not statistically significant and was less than interannual and pasture-level variation in vegetation suggesting that SGI grazing management has negligible effects on vegetation in this system. Similarly, based on linear mixed effect models, grazing metrics tended to have smaller effects on rangeland biomass and productivity compared to environmental factors. For sage-grouse vital rates, our results suggest that annual variation affects hen and chick survival more than grazing management. Nest success did not differ between grazing categories. Future work will

refine hen and chick survival analyses, explore grazing effects on hen and chick seasonal habitat selection, generate population models incorporating sage-grouse vital rates and habitat covariates, link lek counts to demographic rates, and map high quality seasonal sage-grouse habitat.

A related study objective to improve the predictive spatial model of invertebrate biomass has also been completed during this reporting period. Additional models were developed and tested for improving invertebrate biomass predictions across the sage-grouse and songbird study areas. A variety of climate and topographic predictor variables were developed and introduced into the models to assess the potential for improving prediction performance within and across years. Additional grazing metrics (SGI enrollment and grazing timing, frequency, and duration) were tested for importance in previous models (Mitchell et al. 2021), but these variables failed to increase predictive power and were not retained in models detailed within this report. Covariates in the most parsimonious model included: April snowpack, precipitation coefficient of variation, max temperature standard deviation, max temperature coefficient of variation, and accumulated degree days. Despite intensive biomass sampling efforts in the field, prediction improvement was limited. Additional model testing that was conducted rejected the hypothesis that advanced machine learning algorithms could compensate for the strong influence of multi-scale temporal variability on prediction performance over relatively large study areas. Future work will concentrate on appropriately linking information from the invertebrate spatial layers to sage-grouse and songbird populations.

Overall, the effects of SGI rest-rotation grazing were similar to non-SGI grazing regimes for most of the metrics examined. The songbird and invertebrate study objectives are complete with a few remaining deliverables (see further details below). Sage-grouse objectives and results are in progress. We report preliminary results and projected status of completion. Final deliverables for each of the songbird, sage-grouse, and invertebrate projects are expected to be completed by June 2024.

INTRODUCTION

The sagebrush steppe once covered over 62 million hectares in the western US and southwestern Canada but is now among the most imperiled ecosystems in North America (Noss et al. 1995). Conifer encroachment (Miller et al. 2011), exotic annual grass invasion (Chambers et al. 2014; D'Antonio and Vitousek 1992), altered fire regimes (Baker 2011), cropland conversion (Smith et al. 2016), and energy development (Walker et al. 2007; Walston et al. 2009) all contribute to the highly fragmented and disappearing sagebrush biome (Davies et al. 2011; Knick et al. 2003). Sagebrush habitat loss and degradation increase the risk of local and regional extirpations of sagebrush-dependent wildlife, the consequences of which are currently transpiring via emphatic avifaunal declines. Most notably, the greater sage-grouse (*Centrocercus urophasianus*; hereafter 'sage-grouse') has been extirpated from approximately half of its range since European settlement (Schroeder et al. 2004). Long

term declines in sage-grouse abundance and distribution (Connelly and Braun 1997; Schroeder et al. 2004) have warranted multiple evaluations for listing under the Endangered Species Act (U.S. Fish & Wildlife Service, 2015). Similarly, grassland and shrubland songbirds are exhibiting precipitous declines, more so than any other avian guild in North America (Rosenberg et al. 2019; Sauer et al. 2017), with many of these species associated with sagebrush habitat. Rangeland management practices that conserve and improve remnant sagebrush habitats may be a promising strategy for mitigating widespread population declines of sagebrush birds.

Affecting 70% of land in the western US, livestock grazing is the dominant land management practice in sagebrush ecosystems (Heady et al. 1974). While overgrazing has been implicated in sagebrush deterioration (Fleischner 1994; Mack 1981), range conditions have since improved due to advancements in rangeland ecology and better administration of public lands grazing (Holechek 2011). Moreover, grazing is not a novel process in this ecosystem (Perryman et al. 2021); sagebrush steppe associated birds coevolved with variable vegetation structure created by dynamic disturbance processes, including grazing (Duchardt et al. 2018). In some systems, livestock grazing has even been promoted as a surrogate for historical ecological processes (Fuhlendorf and Engle 2001). Since changes to vegetation structure may be the primary mechanism through which grazing affects wildlife, managers may be able to manipulate grazing variables (e.g. stocking rate, timing, duration) to achieve specific habitat outcomes. However, it is difficult to predict effects of grazing prescriptions because effects vary substantially based on ecological setting (Davis et al. 2020).

To combat threats to the sagebrush biome, the Natural Resources Conservation Service (NRCS) - Sage Grouse Initiative (SGI) launched a rest-rotation grazing program designed to simultaneously support wildlife habitat and sustainable ranching (NRCS 2015). The SGI grazing system rotates livestock through different pastures for short periods (< 45 days) and shifts the annual timing of grazing each year. This method is intended to allow vegetation to recover from previous grazing disturbances (NRCS 2017). In contrast, other (i.e. non-SGI) grazing practices may entail a season-long livestock presence without annual changes in season of use (Holecheck et al. 1999). Rest-rotation grazing systems may also promote a mosaic of varying stages of disturbance where the resulting structural heterogeneity fulfills habitat requirements of numerous species (Fuhlendorf and Engle 2001; Krausman et al. 2009). However, grazing impacts are site specific. Intensity, duration, timing, livestock type, and biophysical factors (e.g. soil, climate, topography; Briske et al. 2008; Holechek et al. 1999; Lipsey and Naugle 2017; Veblen et al. 2015) all influence vegetation response to grazing. Additionally, the effects of rest-rotation grazing systems have not been closely examined in central Montana where there is a need for understanding how grazing can meet desired stakeholder and wildlife management goals. Herein, we compare SGI and non-SGI grazing regimes to determine grazing impacts on the sagebrush community within a working landscape.

Management actions applied within the sagebrush steppe are often evaluated through the lens of sage-grouse conservation objectives. Over 50 years of sage-grouse

population declines prompted unprecedented conservation efforts directed towards reversing these trends and precluding protections from the Endangered Species Act. Sage-grouse have subsequently become emblematic of sagebrush conservation and are often perceived as an umbrella species for other sagebrush-dependent species (Barlow et al. 2020; Rowland et al. 2006). While single-species approaches to conservation, like that of sage-grouse, allow managers to make the most of limited resources, this umbrella may not always be adequate for co-occurring species that require separate management actions (Carlisle et al. 2018; Carlisle et al. 2020; Dinkins and Beck 2019; Smith et al. 2021). Alternatively, a multi-species strategy can broaden the protections offered by a single umbrella species since an umbrella species may be unaffected by ecological factors that inevitably limit some co-occurring species (Roberge and Anglestam 2004; Timmer et al. 2019). Systematically selecting multiple focal species that require a range of habitat types and landscape attributes across different spatial scales can provide a holistic perspective on ecosystem integrity and management impacts.

This report uses a multi-species assemblage to examine impacts of rest-rotation grazing employed through SGI. Our focal species included songbirds, sage-grouse, and invertebrates (e.g., arthropods) because of their interrelated roles in sagebrush systems. Songbirds are integral to ecological communities because they function as predators, prey, pollinators, and seed dispersers (Whelan et al. 2008). Their sensitivity to habitat change makes songbirds effective indicators of shifting habitat conditions that may occur as a result of grazing (Canterbury et al. 2000; Coppedge et al. 2006; Milchunas et al. 1998). Sage-grouse conservation has strongly shaped land use policy and management actions that affect other sagebrush-dependent wildlife, but there remains a paucity of information regarding specific grazing effects on sage-grouse demographics and habitat in central Montana (Dettenmaier et al. 2017). Finally, arthropod communities are an important food source for sagebrush songbirds and are especially vital for sage-grouse chick development and survival (Johnson and Boyce 1990).

This annual report outlines the status of a decade (2011-2020) of research evaluating SGI rest-rotation grazing on songbird community and reproduction metrics, sage-grouse habitat and demographics, and invertebrate biomass within a central Montana sagebrush ecosystem. This research has the following long-term objectives:

Songbird

1. Investigate migratory songbird abundance, species richness, species diversity, and community composition responses to SGI versus Non-SGI grazing.
2. Investigate migratory songbird breeding performance of three focal songbird species (Brewer's sparrow, vesper sparrow, and thick-billed longspur) responses to SGI and Non-SGI grazing as a management tool.

Sage-grouse

3. Measure the vegetation response in pastures receiving different grazing and

resting treatments, relative to published sage-grouse habitat needs.

4. Create habitat-based measures of fitness which can be compared among grazing treatments by measuring individual vital rates known to impact population growth in sage-grouse and relating estimated vital rates directly to habitat variables and other important drivers.
5. Identify movements by sage-grouse between grazed and rested pastures to quantify use of treatments proportional to habitat availability and other drivers of sage-grouse resource selection.
6. Create a habitat-linked population model to:
 - a. Evaluate and forecast the benefits of treatments within a rotational grazing system on sage-grouse populations in the context of other drivers of sage-grouse vital rates, so as to put the influence of grazing management on population dynamics in context, and
 - b. Identify current areas that are most important to sage-grouse to prioritize locations where habitat management will have the most benefit to populations.
7. Quantify the population-level response of grazing treatments by indexing lek counts to our population modeling results, then by comparing lek counts within the Roundup study area to surrounding populations. To the extent that lek counts represent population changes reflected in population models, bird response to grazing might be forecasted in other areas where only lek count data are available.
8. Generate spatially-explicit maps for areas with high quality seasonal habitat. Specifically, we will produce maps that delineate areas with habitat attributes that define relative probability of use and that have a positive influence on vital rates during the nesting, brood-rearing, and winter periods, and extrapolate to similar landscapes to the extent that these models validate well.

Invertebrates

9. Improve the predictive spatial model of invertebrate biomass across the sage-grouse and songbird study areas.
10. Link information from the invertebrate spatial layer to sage-grouse and songbird populations,

We have successfully completed 10 years of data collection towards these objectives. Previous years' work is detailed in prior annual reports (see Dreitz et al. 2021; Berkeley et al. 2021; Mitchell et al. 2021). Songbird and invertebrate objectives have been completed with a few final deliverables in progress. Sage-grouse objectives listed above are in progress and we will report on the preliminary results that have been completed thus far. Progress towards objectives, status of deliverables, and future goals are aggregated by project (songbird, sage-grouse, invertebrates) within this report.

STUDY AREA

The study area was in central Montana in rolling topography that ranged from 975-1,250m in elevation (Smith et al. 2018b) and covered approximately 150,000 hectares in Musselshell and Golden Valley counties (Figure 1). The vegetation was consistent with big sagebrush steppe, the most widely distributed sagebrush system in Montana. Wyoming big sagebrush (*Artemisia tridentata ssp. wyomingensis*) and silver sagebrush (*A. cana*) were both common and co-dominant, with a mix of perennial bunchgrasses, perennial rhizomatous grasses, and forbs composing up to 25% of cover (Montana Natural Heritage Program 2021). This region has cooler soil temperature and higher soil moisture than other parts of the sage-grouse range (Pyke et al. 2015). The average monthly temperature in Roundup (2009-2020) ranged from a low of -3.8° Celsius (25.1° Fahrenheit) in January to a high of 21.8° Celsius (71.2° Fahrenheit) in July (National Centers for Environmental Information 2021). Average monthly precipitation in Roundup (2009-2020) ranged from a low of 9.40 millimeters (0.37 inches) in January to a high of 73.41 millimeters (2.89 inches) in June (National Centers for Environmental Information 2021). The climate is cold semi-arid (Pyke et al. 2015), with distinct seasons that include cool and wet springs, hot and dry summers, cool and wet autumns, and cold, snowy winters. The study area is a mosaic of public (federal, state, and county) and private ownership dominated by cattle rangeland, with some sheep rangeland and some dryland farming (Smith et al. 2018a, Smith et al. 2018b).

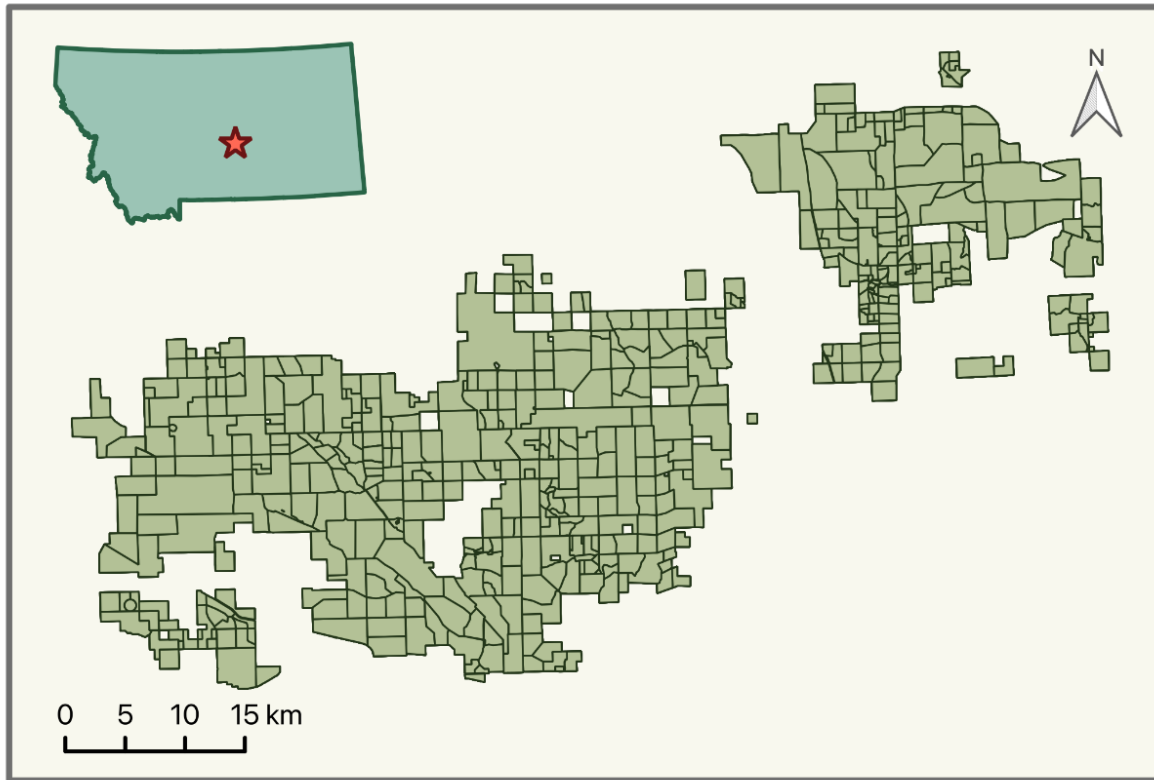


Figure 1: Livestock pasture boundaries within the study area in Golden Valley and Musselshell Counties, Montana, USA during 2011-2020.

THE SAGE GROUSE INITIATIVE (SGI) GRAZING PROGRAM

The SGI grazing program in central Montana focused on improving livestock production and rangeland health while simultaneously alleviating threats to and improving habitat for greater sage-grouse (NRCS 2015). The SGI program was implemented on private ranches containing potential sage-grouse habitat as defined by topography and sagebrush canopy cover $\geq 5\%$ (NRCS pers. comm.) within sage-grouse core areas (Figure 2). FWP has designated core areas in Montana as locations of highest conservation value for sage-grouse based on habitat and number of breeding males. FWP has estimated that the core areas included $\sim 76\%$ of the displaying males in Montana as of 2013.

Livestock producers enrolled in the SGI program implemented an approximately three-year grazing regime developed with NRCS range management specialists. SGI grazing regimes were rotational and used a combination of rest and deferment to increase vegetation cover for nesting hens (Doherty et al. 2014, Smith et al. 2018b), in addition to other strategies. Range management specialists suggested pasture rest, pasture

deferment, changed the number of animal units, or installed fences or water sources to adjust pasture size or livestock distribution. SGI grazing regimes were tailored to each ranch and varied by needs of the producer or pasture condition while following the NRCS Conservation Practice Standard for Prescribed Grazing (Natural Resources Conservation Service 2017, Smith et al. 2018b). Additionally, plans align with four minimum criteria intended to support sage-grouse habitat:

1. Grazing utilization rates of $\leq 50\%$ of the current year's key forage species growth,
2. ≥ 20 -day shift annually in the timing of grazing,
3. A plan to address unexpected circumstances like drought or fire, and
4. ≤ 45 -day continuous grazing within any one pasture (Smith et al. 2018b).

Our work evaluated the effects of these recommendations to determine if this program yields biologically-relevant benefits to songbirds, sage-grouse, and their invertebrate food sources. We categorized enrolled pastures into before, during, and after implementation of SGI grazing to disentangle direct and indirect effects of SGI grazing. Non-SGI grazing involved multiple types of grazing systems with less intensively managed and slower rotations, usually lacking annual changes in use.

SONGBIRD COMMUNITY AND REPRODUCTION

From 2013-2019 we collected field data to evaluate the relationships between grazing and sagebrush steppe songbird community composition and demographic parameters related to SGI's rotational grazing regime. Detailed methods and results can be found in the Migratory Songbird Grazing Study Final Report (P-R grant W-165-R-1 to FWP; Dreitz et al. 2021). This study had two objectives, which are detailed below.

Objective 1: Investigate migratory songbird abundance, species richness, species diversity, and community composition responses to SGI versus Non-SGI grazing.

We conducted avian count transect surveys using the dependent double-observer method. During 2013–2019, the total number of individuals we observed in the study area, regardless of grazing regime, ranged from 5,954–14,097, and the total number of species ranged from 72–88. We observed low variation in avian community composition amongst years, suggesting a relatively stable species richness in our study area over time. The migratory songbird species observed most often since 2013 were: thick-billed longspur (*Rhynchophanes mccownii*; previously named McCown's longspur), vesper sparrow (*Pooecetes gramineus*), Brewer's sparrow (*Spizella breweri*), horned lark (*Eremophila alpestris*), and western meadowlark (*Sturnella neglecta*). Estimates of abundance for the five most common species suggest species-specific responses to grazing (see Dreitz et al. 2021 for specific categorizations of SGI and non-SGI pastures used in songbird community analyses). For instance, thick-billed longspurs were most

abundant on lands using SGI grazing, while observations of western meadowlarks were higher on non-SGI plots during the early years of our study.

Objective 2: Investigate migratory songbird breeding performance of three focal songbird species responses to SGI and Non-SGI grazing as a management tool.

We identified three focal species, each associated with one of the three most common vegetation characteristics in sagebrush (*Artemisia* spp.) steppe. We conducted nest searches and monitored nesting activity of Brewer’s sparrow (sagebrush nester), vesper sparrow (generalist ground nester), and thick-billed longspur (grassland ground nester). For our three focal species, we located 40% of nests on lands using non-SGI grazing, compared to 60% of nests on lands using SGI grazing. Estimated nest density was higher on SGI grazing plots for thick-billed longspur while nest densities for both Brewer’s and vesper sparrows were higher on non-SGI plots. For all focal species, nest success showed little difference between SGI and non-SGI (see Dreitz et al. 2021 for specific categorizations of SGI and non-SGI pastures used in songbird reproduction analyses). Both nest density and nest success varied annually.

Future Goals:

The PhD student is currently refining analyses and writing dissertation chapters related to 1) further assessment of linking field grazing metrics to remotely sensed biomass and productivity, 2) using songbird nest success field data to estimate nest density and nest success by development and testing of a novel model, and 3) implementation of the novel model to determine what factors are most influential to sagebrush steppe songbird reproduction. The individual is on track to complete dissertation work and earn their degree during the fall 2022 semester.

Deliverables:

Objectives	Description	Status
Objective 1	Investigate migratory songbird abundance, species richness, species diversity, and community composition responses to SGI versus Non-SGI grazing.	Complete
Objective 2	Investigate migratory songbird breeding performance of three focal songbird species responses to SGI and Non-SGI grazing as a management tool.	Complete

Student	Description	Status
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MS Student	Individual officially enrolled at the University of Montana in August 2016 (year 1). Student participated in 2016, 2017, and 2018 field seasons, year 1, 2, and 3 respectively	Completed: Degree awarded May 2019
PhD Student	Selected PhD candidate in spring 2017 (year 2). Individual officially enrolled at the University of Montana in August 2017 and is on track to achieve their graduate degree in fall 2022.	In Progress: Expected completion fall 2022

Publications	Status
Ruth, K. A., L. I. Berkeley, K. M. Strickfaden, and V. J. Dreitz. In Review. Density dependence of songbird demographics in grazed sagebrush steppe. PLoS ONE.	In Review
Reintsma, K.M., A.H. Harrington, V.J. Dreitz. 2019. Validation of a novel time-to-event nest density estimator on passerines: An example using Brewer’s sparrows (<i>Spizella breweri</i>). PLoS ONE 12:e0227092	Complete
Reintsma, K. M., V. J. Dreitz, L. I. Berkeley. 2022. Thick-billed Longspur (<i>Rhynchophanes mccownii</i>) reproduction shows minimal short-term response to conservation-based program. <i>Wilson Journal of Ornithology</i> . 134 (2): 365–372	Complete
Golding, J.D., J.J. Nowak, and V.J. Dreitz. 2017. A multispecies dependent double-observer model: A new method for estimating multispecies abundance. <i>Ecology and Evolution</i> 7:3425–3435.	Complete
Golding, J.D. and V.J. Dreitz. 2017. Songbird response to rest-rotation and season-long cattle grazing in a grassland sagebrush ecosystem. <i>Journal of Environmental Management</i> 204: 605-612.	Complete
J. D. Golding and V. J. Dreitz. 2016. Comparison of removal-based methods for estimating abundance of five species of prairie songbirds. <i>Journal of Field Ornithology</i> 87: 417–426 (PDF)	Complete

GREATER SAGE-GROUSE DEMOGRAPHICS AND HABITAT

We evaluated the effectiveness of SGI rotational grazing systems to manage sage-grouse habitat in central Montana. To do so, we measured sage-grouse vital rates (including nest success, chick survival, and hen survival) within these grazing systems and compared them with vital rates in non-SGI grazing systems. We included factors describing vegetation structure and composition to evaluate the effects of grazing on sage-grouse habitat. Data collection occurred during 2011-2020. Detailed methods and results can be found in the “The Effects of Grazing on Greater Sage-Grouse Population Dynamics and Habitat in Central Montana” report (PR grant W-158-R to FWP; Berkeley et al. 2021). Our preliminary results presented below represent progress on study objectives.

Objective 1: Measure the vegetation response in pastures receiving different grazing and resting treatments, relative to published sage-grouse habitat needs.

During 2012-2019, we measured herbaceous vegetation in potential sage-grouse habitat using the line-intercept technique at a set of random field plots stratified by grazing system (SGI and non-SGI) to test for differences in vegetation metrics across the project area. While effects were not statistically significant, we observed tendencies for total herbaceous vegetation, live grass height, annual perennial forb cover (derived from the Rangeland Analysis Platform, or RAP), and RAP-derived shrub cover to vary with SGI grazing management, and an effect of SGI grazing management on RAP-derived litter cover, indicating that there was more herbaceous vegetation and less forb and shrub cover present in pastures currently being grazed according to the SGI program protocols. The variation among years and pastures was greater than differences observed in these metrics, reflecting the wide variation in vegetation growth. Our results are consistent with preliminary analyses from Smith et al. (2018a) that found negligible effects of SGI grazing management on vegetation in sage-grouse habitat.

Since the completion of the 2021 sage-grouse PR report (see Berkeley et al. 2021), we have continued analyses of vegetation in the study area. We used linear mixed effects models to test for effects of grazing management on rangeland biomass and productivity metrics while accounting for variation across years and pastures. In these models, remotely sensed biomass or productivity were response variables and field-based grazing data combined with remotely sensed abiotic and biotic environmental factors were explanatory variables. We found point-level field measures of grazing (e.g., cow patties, percentage of dung in Daubenmire plots, and number of plants grazed) showed positive effects, especially on perennial plant rangeland responses. Grazing measures at the pasture-level showed a small negative effect on annual plant rangeland responses. Grazing metrics tend to have smaller covariate effects on rangeland biomass and productivity compared to environmental factors, indicating a greater importance of environmental factors in influencing rangeland productivity and biomass in this study area.

Objective 2: Create habitat-based measures of fitness which can be compared among grazing treatments by measuring individual vital rates known to impact population growth in sage-grouse and relating estimated vital rates directly to habitat variables and other important drivers.

We collected nest data from 2011-2020 on sage-grouse hens, including yearling (second-year) females and adult (after-second-year) females. Sage-grouse nests were located by monitoring pre-nesting females using radio telemetry twice per week at the onset of the nesting season. We estimated sage-grouse nest daily survival rate (DSR) accounting for exposure days (e.g., days between field observations) and derived an estimate of nest success. The SGI status and grazing treatment of each nest location was used to investigate livestock grazing effects on estimated DSR. We evaluated 11

competing models containing grazing management, year, and nest-specific variables that we predicted would affect sage-grouse nest success. The top model contained a covariate effect for “year”, while none of the models with other variables (SGI grazing categorization, grazing treatment, or nest attempt) were competitive. Using the top model, nest success of 664 nests across the 37-day nesting period averaged 0.36 (95% CI = 0.31-0.40) from 2011-2019. There were no biologically meaningful differences in nest success across SGI grazing categories. Overall, these results suggest that the year-to-year variation experienced by nesting sage-grouse is greater than any direct or indirect effect of SGI grazing management.

Chicks were monitored using VHF radio transmitters that were attached to four randomly selected chicks per brood. We used a Kaplan-Meier survival function to evaluate chick survival with staggered entry designs and right-censoring for individuals with unknown fates, dropped transmitters, or that survived until their transmitters expired. We used a Cox proportional hazards model to assess the effects of time-varying and continuous covariates on chick mortality risk including body condition index of the brood hen; chick mass; sexID of chicks; weather; vegetation cover metrics extracted from the Rangeland Analysis Platform (RAP); other remotely sensed variables including slope, compound topographic index, vector ruggedness measure, herbaceous vegetation heights, shrub heights; distance to crops or roads; remotely sensed anthropogenic disturbance metrics including human structure density and human disturbance index; and grazing metrics. Annual survival estimates for 521 chicks radio marked during 2011-2019 to 75 d post-hatch (when they are the size of adults) differed among years and ranged from 0.19-0.60. The median survival time for chicks with data from all years pooled was 42 days (95% confidence interval [CI] = 33-59 d). None of the remote-sensed vegetation metrics obtained from RAP had a significant effect on the mortality risk of chicks. There was also not a statistically significant difference between female and male chick survival, though there was a tendency for females to have a higher probability of survival than males. Our initial analyses also suggested that the probability of chick mortality was higher when using pastures during/post SGI grazing management. However, small sample sizes for chick locations may have given a spurious result given that we had limited power to detect differences.

We captured hens at the start of the breeding season (March and April) from 2011-2019. Hens were marked with VHF transmitters or GPS PTTs that we used to monitor hen location and survival. For this report, we focus on the VHF hen locations and will include GPS PTT data in the final report. We used a Kaplan-Meier survival function to evaluate hen survival with staggered entry designs and right-censoring for individuals with unknown fates, dropped transmitters, or that survived until their transmitters expired. The SGI status and grazing treatment of each hen location was used to investigate livestock grazing effects on hen survival. The median survival time estimated for 495 marked hens monitored from 2011-2020 was 1.25 years (95% CI = 1.16-1.40 yrs). Preliminary results suggest that both year-to-year variation and SGI grazing have the potential to affect hen survival, though differences due to grazing effects were minor and likely not significant enough to change current management. Overall, our results

suggest that annual variation has more of an effect on sage-grouse vital rates than grazing management.

Future Goals:

Remaining products will be delivered by the end of our current PR Grant # F21AF01330. These deliverables include 1) refining hen and chick survival analyses, 2) exploring the effects of grazing on seasonal habitat selection of hens and chicks and population dynamics, 3) producing a population model that incorporates sage-grouse vital rates and habitat covariates and assesses the effects of grazing at the population level, 4) exploring the link between lek counts and demographic rates, and 5) generating a spatial map of high-quality seasonal habitat.

For ongoing analyses of nest success and survival of hens and chicks, we plan to include more covariates that will encompass some of the variation that is currently attributed to annual variation. These variables include weather metrics (e.g. temperature and precipitation) and fine-scale vegetation metrics (e.g. shrub, grass, and forb cover). We aim to have a chick survival manuscript submitted by December 2022. Additionally, we are in the process of analyzing habitat use of hens and chicks. We will determine if habitat selection aligns with areas of high or low hen or chick survival, and how this is affected by SGI grazing management. The results from Objective #2 (e.g., vital rates of nest success, chick survival, and hen survival) will be included in population models to complete objectives 4-6,

The PhD student is currently refining analyses and writing dissertation chapters related to hen habitat selection, nest success, and population model comparisons. The individual is on track to complete dissertation work and earn their degree during the fall 2022 semester.

Deliverables:

Objective	Description	Status
Objective 1	Measure the vegetation response in pastures receiving different grazing and resting treatments, relative to published sage-grouse habitat needs	Complete
Objective 2	Create habitat-based measures of fitness which can be compared among grazing treatments by measuring individual vital rates known to impact population growth in sage-grouse and relating estimated vital rates directly to habitat variables and other important drivers	In Progress
Objective 3	Identify movements by sage-grouse between grazed and rested pastures to quantify use of treatments proportional to habitat availability and other drivers of sage grouse resource selection	In Progress

Objective 4	Create a habitat-linked population model to: 1) Evaluate and forecast the benefits of treatments within a rotational grazing system on sage-grouse populations in the context of other drivers of sage grouse vital rates, so as to put the influence of grazing management on population dynamics in context, and 2) Identify current areas that are most important to sage grouse to prioritize locations where habitat management will have the most benefit to populations	In Progress
Objective 5	Quantify the population-level response of grazing treatments by indexing lek counts to our population modeling results, then by comparing lek counts within the Roundup study area to surrounding populations. To the extent that lek counts represent population changes reflected in population models, bird response to grazing might be forecasted in other areas where only lek count data are available	In Progress
Objective 6	Generate spatially-explicit maps for areas with high quality seasonal habitat. Specifically we will produce maps that delineate areas with habitat attributes that define relative probability of use and that have a positive influence on vital rates during the nesting, brood-rearing, and winter periods, and extrapolate to similar landscapes to the extent that these models validate well	In Progress

Student	Description	Status
PhD Student	Selected PhD candidate in fall 2018. Individual officially enrolled at the University of Montana in August 2018 and is on track to achieve their graduate degree in fall 2022.	In Progress: Expected completion fall 2022

Publications	Status
Hen Survival Manuscript (Objective 2)	In Progress: Expected to submit by Mar 2023
Habitat Selection Manuscript (Objectives 3 & 6)	In Progress: Expected to submit by Mar 2023
Nest Success Manuscript (Objective 2)	In Progress: Expected to submit by Jan 2023
Chick Survival Manuscript (Objective 2)	In Progress: Expected to submit by Dec 2022
Habitat-linked Population Model Manuscript (Objective 4)	In Progress: Expected to submit by Jan 2024
Lek Counts Manuscript (Objective 5)	In Progress: Expected to submit by May 2023

Smith, J. T., J. D. Tack, L. I. Berkeley, M. Szczypinski, and D. E. Naugle. 2018a. Effects of Rotational Grazing Management on Nesting Sage-Grouse. <i>Journal of Wildlife Management</i> , 82:103-112. doi:10.1002/jwmg.21344.	Complete
Smith, J. T., J. D. Tack, L. I. Berkeley, M. Szczypinski, and D. E. Naugle. 2018b. Effects of Livestock Grazing on Nesting Sage-Grouse in Central Montana. <i>Journal of Wildlife Management</i> , 82:1503-1515. doi:10.1002/jwmg.21500.	Complete
Smith, J. T., J. D. Tack, K. E. Doherty, B. W. Allred, J. D. Maestas, L. I. Berkeley, S. Dettenmaier, T. A. Messmer, D. E. Naugle. 2017. Phenology largely explains taller grass at successful nests in greater sage-grouse. <i>Ecology and Evolution</i> , 8:356-364. doi: 10.1002/ece3.3679.	Complete

INVERTEBRATE BIOMASS PREDICTIVE SPATIAL LAYER

The objective of this project was to create a predictive spatial layer of invertebrate biomass across the sage-grouse (PR grant #F15AF00490 “MT Sage-Grouse Grazing Evaluation”) and songbird (PR grant #F16AF00294 “Migratory Songbird Grazing Study”) grazing project study areas in central Montana to provide invertebrate food availability data for sage-grouse grazing project vital rate, habitat use, and population models, and songbird grazing project reproduction, community, and abundance models. We completed data collection during spring/summer 2020 and generated a predictive invertebrate biomass spatial layer in 2021 for the sage-grouse and songbird study areas (Mitchell et al. 2021). Results reported herein represent further analysis and fine-tuning of the invertebrate biomass spatial layer for PR grant # F21AF01330. Detailed methods and results can be found in “Predictive Spatial Layer of Invertebrate Biomass for Sage-Grouse and Songbird Grazing Studies in Central Montana” report (PR grant W-164-R-1 to FWP; Mitchell et al. 2021).

Objective 1: Create a spatial layer that predicts invertebrate biomass for the sage-grouse and songbird grazing project study areas.

Building on previous work (see Mitchell et al. 2021), we focused on identifying meaningful predictors of arthropod biomass that fit the sampling strategy employed during the 2019 and 2020 field seasons. While arthropod biomass data were collected for 2012-2020, we constrained observations to 2019 and 2020 because these years had identical sampling frameworks. Previous years used both pitfall and sweep-net samples, whereas 2019 and 2020 used only sweep nets. We represented each sampling location as a point feature and extracted variables related to cumulative growing degree-days, soil moisture, extreme weather, and late spring snow since these are important drivers of arthropod growth (Shaftel et al. 2021; Telfer and Hassall 1999; Wu et al. 2021). These variables were used to compare the predictive power of temporally-static versus dynamic variables. For the biomass data, we aggregated individual survey events by location and date/time to get total biomass for each site. Biomass was then log-transformed. To standardize interpretation of effect sizes, the log-transformed biomass plus all other numeric variables were scaled to have a mean and standard deviation values of 0 and 1, respectively. We used these data to predict log biomass using spatiotemporal variables in ordinary-least-squares (OLS) and linear mixed effects models.

We produced four models to explain the variance of arthropod biomass and used Akaike Information Criterion (AIC) (Akaike 1991) to measure bias and select the most parsimonious model. The first model was a linear mixed effects model containing 20 predictors (Table 1; Table 2) and fixed effects for each year. The conditional R^2 was 0.66, Intra-class Correlation was 0.54, and the AIC value was 996.33. While this model had decent predictive power and reasonable within-year correlation, it had a high degree of bias, and many of the predictors were insignificant. The second model was a mixed effects model with annual fixed effects. This model contained topographic

variables (slope; Topographic Wetness Index (TWI); elevation) and the most significant predictors from the first mixed effects model, including the coefficient of variation of precipitation¹, standard deviation of daily max temperature¹, coefficient of variation of daily max temperature¹, total precipitation¹, and cumulative degree days² (with a 17.8 degree C threshold) (Brust, 2009). A conditional R² of 0.63, Intra-class Correlation of 0.47, and lower AIC of 955.9 indicates a more parsimonious model, but there is still a high degree of bias. To reduce bias, we fit two OLS models. To compensate for the observed intra-class correlation in the mixed models that could not be replicated in an OLS model, we incorporated total April snowfall³ as a predictor. This approach theoretically provided some of the information lost in the annual effects since April snowfall differed significantly between 2019 and 2020 ($p < 0.0001$). The first OLS model had the same predictors as the previous mixed effects model and produced an adjusted R² of 0.64, and a lower AIC of 940.9. For the second OLS model, we removed redundant topographic predictors. This final model yielded an adjusted R² of 0.65 and the lowest AIC value of 928, indicating the most parsimonious model yet using only weather-related covariates (Table 2).

Given the large spatial extent and lack of repeat site visits, we were unable to disentangle individual observations from the temporal window in which they were taken. This is due to mechanistic relationships between weather and arthropod life stages, coupled with a sample size of 1 for each unique site (Brust et al. 2009; Shaftel et al. 2021). Previous studies that have attempted to predict arthropod biomass with remote sensing have used drones, as the timing of imagery acquisition must be nearly identical to that of the field sampling for reliable results (Traba et al. 2022). Results from this study indicate that future sampling efforts should include static plots with repeat sampling at regular intervals to reduce temporal confounding of spatial covariates.

¹ Variable was calculated for the 60 days prior to the sample date, roughly equal to the average lifespan of the most abundant arthropods (Orthoptera).

² Variable was calculated from the start of the calendar year.

³ Variable was extracted for a single month in the year of sampling.

Table 1: Climate, topographic, and field data predictor variables used in invertebrate model development.

Climate Predictors	Source	Spatial Resolution	Temporal Resolution
Average maximum June temperature between 2011 and 2020	Daymet (Thornton et al. 2020)	1 km	1 day
Average maximum July temperature between 2011 and 2020	Daymet	1 km	1 day
Average maximum August temperature between 2011 and 2020	Daymet	1 km	1 day
Average maximum June precipitation (water equivalent) between 2011 and 2020	Daymet	1 km	1 day
Average maximum July precipitation (water equivalent) between 2011 and 2020	Daymet	1 km	1 day
Average maximum August precipitation (water equivalent) between 2011 and 2020	Daymet	1 km	1 day
Total precipitation in water equivalency during 60 days prior to sampling	Daymet	1 km	1 day
Cumulative degree days: sum of differences between temperature threshold (17 C) and daily average temperature	Daymet	1 km	1 day
Daily max temperature coefficient of variation 60 days prior to sampling	Daymet	1 km	1 day
Daily precipitation coefficient of variation 60 days prior to sampling	Daymet	1 km	1 day
Number of days within 60 days before the survey where the temperature dropped one standard deviation below the mean temperature	Daymet	1 km	1 day
Number of days within 60 days before the survey where the max daily temperature increased one standard deviation above the mean temperature	Daymet	1 km	1 day
Daily max temperature standard deviation 60 days prior to sampling	Daymet	1 km	1 day
Daily precipitation standard deviation 60 days prior to sampling	Daymet	1 km	1 day
Average total snow water equivalent (SWE) between 2011 and 2020	SNODAS (NOHRSC 2004)	1 km	1 day
April SWE of the sample year	SNODAS	1 km	1 day
Topographic Predictors			
Topographic Wetness Index	NED (USGS 2012)	10 m	NA
Slope	NED	10 m	NA

Elevation	NED	10 m	NA
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Other Predictors

Year	NA	NA	NA
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Table 2: Model coefficients and their estimates for the final ordinary-least-squares model predicting invertebrate biomass.

Predictors	Estimates	Confidence Intervals	<i>p</i> -value
Intercept	0	-0.05 – 0.05	1
April Snow Water Equivalent	0.3	0.24 – 0.37	<0.001
Precipitation Coefficient of Variation	0.07	0.00 – 0.14	0.036
Max Temperature Standard Deviation	0.23	0.14 – 0.32	<0.001
Max Temperature Coefficient of Variation	-0.44	-0.56 – -0.32	<0.001
Cumulative Degree-Days	0.35	0.25 – 0.45	<0.001
Observations	512		
R2 / R2 adjusted	0.644 / 0.640		

Future Goals:

We have completed exploring the ability to make a predictive spatial layer for invertebrate biomass. Future work will qualitatively link invertebrate biomass model predictions to the songbird and sage-grouse vital rate and population models.

Deliverables:

Objectives	Description	Status
Objective 1	Create a spatial layer that predicts invertebrate biomass for the sage-grouse and songbird grazing project study areas.	Complete

Publications	Status
Goosey, H. B., J. T. Smith, K. M. O'Neill, and D. E. Naugle. 2019. Ground-Dwelling Arthropod Community Response to Livestock Grazing: Implications for Avian Conservation. <i>Environmental Entomology</i> 48:856-866.	Complete

LITERATURE CITED

- Akaike H. 1991. Information theory and an extension of the maximum likelihood principle. In: Kotz S, Johnson NL, editors. *Breakthroughs in statistics*. New York, NY: Springer New York. (Springer Series in Statistics). p. 610-624.
- Baker WL. 2011. Pre-Euro-American and recent fire in sagebrush ecosystems. In: Knick ST, Connelly JW, editors. *Greater sage-grouse ecology and conservation of a landscape species and its habitats*. Berkeley: University of California Press. (Studies in Avian Biology; no. 38). p. 185–201.
- Barlow NL, Kirol CP, Doherty KE, Fedy BC. 2020. Evaluation of the umbrella species concept at fine spatial scales. *J Wildl Manage*. 84(2):237–248.
- Berkeley L, Szczypinski M, Helm J, Dreitz VJ. 2021. The effects of grazing on greater sage-grouse population dynamics and habitat in central Montana. *Montana Fish, Wildlife and Parks*.
https://fwp.mt.gov/binaries/content/assets/fwp/conservation/sage-grouse/pr_finalreport_duedec2021_v20_final.pdf.
- Briske DD, Derner JD, Brown JR, Fuhlendorf SD, Teague WR, Havstad KM, Gillen RL, Ash AJ, Willms WD. 2008. Rotational grazing on rangelands: reconciliation of perception and experimental evidence. *Rangeland Ecol Manage*. 61(1):3–17.
- Brust ML, Hoback WW, Wright RJ. 2009. Degree-day requirements for eight economically important grasshoppers (Orthoptera: Acrididae) in Nebraska using field data. *Environ Entomol*. 38(5):1521–1526.
- Canterbury GE, Martin TE, Petit DR, Petit LJ, Bradford DF. 2000. Bird communities and habitat as ecological indicators of forest condition in regional monitoring. *Conserv Biol*. 14(2):544–558.
- Carlisle JD, Chalfoun AD. 2020. The abundance of greater sage-grouse as a proxy for the abundance of sagebrush-associated songbirds in Wyoming, USA. *Avian Conserv Ecol/Ecol Conserv Oiseaux*. 15(2).
- Carlisle JD, Keinath DA, Albeke SE, Chalfoun AD. 2018. Identifying holes in the greater sage-grouse conservation umbrella. *J Wildl Manage*. 82(5):948–957.
- Chambers JC, Miller RF, Board, David I., Pyke DA, Roundy BA, Grace JB, Schupp EW, Tausch RJ. 2014. Resilience and resistance of sagebrush ecosystems: implications for state and transition models and management treatments. *Rangeland Ecol Manage*. 67(5):440–454.

- Connelly JW, Braun CE. 1997. Long-term changes in sage grouse *Centrocercus urophasianus* populations in western North America. *Wildlife Bio.* 3(3/4):229–234.
- Coppedge BR, Engle DM, Masters RE, Gregory MS. 2006. Development of a grassland integrity index based on breeding bird assemblages. *Environ Monit Assess.* 118(1-3):125–145.
- D'Antonio CM, Vitousek PM. 1992. Biological invasions by exotic grasses, the grass/fire cycle, and global change. *Annual Review of Ecology and Systematics.* 23:63–87.
- Davies KW, Boyd CS, Beck JL, Bates JD, Svejcar TJ, Gregg MA. 2011. Saving the sagebrush sea: an ecosystem conservation plan for big sagebrush plant communities. *Biol Conserv.* 144(11):2573–2584.
- Davis KP, Augustine DJ, Monroe AP, Derner JD, Aldridge CL. 2020. Adaptive rangeland management benefits grassland birds utilizing opposing vegetation structure in the shortgrass steppe. *Ecol Appl.* 30(1):e02020.
- Dettenmaier SJ, Messmer TA, Hovick TJ, Dahlgren DK. 2017. Effects of livestock grazing on rangeland biodiversity: a meta-analysis of grouse populations. *Ecol Evol.* 7(19):7620–7627.
- Dinkins JB, Beck JL. 2019. Comparison of conservation policy benefits for an umbrella and related sagebrush-obligate species. *Human - Wildlife Interactions.* 13(3):447–458.
- Doherty KE, Naugle DE, Tack JD, Walker BL, Graham JM, Beck JL. 2014. Linking conservation actions to demography: grass height explains variation in greater sage-grouse nest survival. *Wildlife Biol.* 20(6):320–325.
- Dreitz VJ, Reintsma K, Delamont M, Berkeley L. 2021. Final report migratory songbird – grazing. Montana Fish, Wildlife and Parks. https://fwp.mt.gov/binaries/content/assets/fwp/conservation/wildlife-reports/nongame/songbirdsgrazinggroundup_2021_final.pdf.
- Duchardt CJ, Porensky LM, Augustine DJ, Beck JL. 2018. Disturbance shapes avian communities on a grassland-sagebrush ecotone. *Ecosphere.* 9(10):e02483.
- Fleischner TL. 1994. Ecological costs of livestock grazing in western North America. *Conserv Biol.* 8(3):629–644.
- Fuhlendorf SD, Engle DM. 2001. Restoring heterogeneity on rangelands: ecosystem management based on evolutionary grazing patterns. *Bioscience.* 51(8):625–632.

- Heady HF, Box TW, Butcher JE, Colbert FT, Cook CW, Eckert RE, Gray JR, Hedrick DW, Hodgson HJ, Kearl WG, et al. 1974. Livestock grazing on federal lands in the 11 western states. *J Range Manage.* 27(3):174.
- Holechek J. 2011. *Range management: principles and practices*. 6th ed. Boston: Prentice Hall.
- Holechek JL, de Souza Gomes H, Molinar F, Galt D. 1999. Grazing studies: what we've learned. *Rangelands.* 21(2):12-16.
- Johnson GD, Boyce MS. 1990. Feeding trials with insects in the diet of sage grouse chicks. *J Wildl Manage.* 54(1):89–91.
- Knick ST, Dobkin DS, Rotenberry JT, Schroeder MA, Vander Haegen WM, van Riper C. 2003. Teetering on the edge or too late? Conservation and research issues for avifauna of sagebrush habitats. *Condor.* 105(4):611–634.
- Krausman PR, Naugle DE, Frisina MR, Northrup R, Bleich VC, Block WM, Wallace MC, Wright JD. 2009. Livestock grazing, wildlife habitat, and rangeland values. *Rangelands.* 31(5):15–19.
- Lipsev MK, Naugle DE. 2017. Precipitation and soil productivity explain effects of grazing on grassland songbirds. *Rangeland Ecol Manage.* 70(3):331–340.
- Mack RN. 1981. Invasion of *Bromus tectorum* L. into western North America: an ecological chronicle. *Agro-Ecosyst.* 7(2):145–165.
- Milchunas DG, Lauenroth WK, Burke IC. 1998. Livestock grazing: animal and plant biodiversity of shortgrass steppe and the relationship to ecosystem function. *Oikos.* 83(1):65–74.
- Miller RF, Knick ST, Pyke DA, Meinke CW, Hanser SE, Wisdom MJ, Hild AL. 2011. Characteristics of sagebrush habitats and limitations to long-term conservation. In: Knick ST, Connelly JW, editors. *Greater sage-grouse ecology and conservation of a landscape species and its habitats*. Berkeley: University of California Press. (Studies in Avian Biology; no. 38). p. 145–184.
- Mitchell J, Tobalske C, Berkeley L, Szczypinski M, Dreitz VJ, Helm J. 2021. Predictive spatial layer of invertebrate biomass for sage-grouse and songbird grazing studies in central Montana. *Montana Fish, Wildlife, and Parks.*
https://fwp.mt.gov/binaries/content/assets/fwp/conservation/sage-grouse/bug_final_pr_report_2021_final2.pdf.
- Montana Natural Heritage Program. 2021. Big sagebrush steppe. *Montana Field Guide.*
http://fieldguide.mt.gov/displayES_Detail.aspx?es=5454. Accessed 30 Nov 2021.

- National Centers for Environmental Information. 2021. US Climate normals quick access. <https://www.ncei.noaa.gov/access/us-climate-normals>. Accessed 1 Nov 2021.
- NOHRSC (National Operational Hydrologic Remote Sensing Center). 2004. Snow Data Assimilation System (SNODAS) Data Products at NSIDC, Version 1. <https://nsidc.org/data/G02158/versions/1>.
- NRCS (Natural Resources Conservation Service). 2015. Outcomes in conservation: Sage Grouse Initiative. U.S. Department of Agriculture. https://www.sagegrouseinitiative.com/wp-content/uploads/2015/02/NRCS_SGI_Report.pdf.
- NRCS (Natural Resources Conservation Service). 2017. NRCS conservation practice standard prescribed grazing. US Department of Agriculture Report No.: 528. https://www.nrcs.usda.gov/Internet/FSE_DOCUMENTS/stelprdb1255132.pdf.
- Noss RF, Laroe ET III, Scott JM. 1995. Endangered ecosystems of the United States: a preliminary assessment of loss and degradation. Washington, D.C., USA: U.S. Department of the Interior, National Biological Service Report No.: Biological Report 28.
- Perryman BL, Schultz BW, Meiman PJ. 2021. Forum: A change in the ecological understanding of rangelands in the Great Basin and Intermountain West and implications for management: revisiting Mack and Thompson (1982). *Rangeland Ecol Manage.* 76:1–11.
- Pyke DA, Chambers JC, Pellant M, Knick ST, Miller RF, Beck JL, Doescher PS, Schupp EW, Roundy BA, Brunson M, et al. 2015. Restoration handbook for sagebrush steppe ecosystems with emphasis on greater sage-grouse habitat—Part 1. Concepts for understanding and applying restoration. Reston, VA: U.S. Geological Survey Circular Report No.: 1416. <http://pubs.er.usgs.gov/publication/cir1416>.
- Roberge J-M, Angelstam P. 2004. Usefulness of the umbrella species concept as a conservation tool. *Conserv Biol.* 18(1):76–85.
- Rosenberg KV, Dokter AM, Blancher PJ, Sauer JR, Smith AC, Smith PA, Stanton JC, Panjabi A, Helft L, Parr M, et al. 2019. Decline of the North American avifauna. *Science.* 366(6461):120–124.
- Rowland MM, Wisdom MJ, Suring LH, Meinke CW. 2006. Greater sage-grouse as an umbrella species for sagebrush-associated vertebrates. *Biol Conserv.* 129(3):323–335.

- Sauer JR, Pardieck KL, Ziolkowski DJ, Smith AC, Hudson M-AR, Rodriguez V, Berlanga H, Niven DK, Link WA. 2017. The first 50 years of the North American Breeding Bird Survey. *Condor*. 119(3):576–593.
- Schroeder MA, Aldridge CL, Apa AD, Bohne JR, Braun CE, Bunnell SD, Connelly JW, Deibert PA, Gardner SC, Hilliard MA, et al. 2004. Distribution of sage-grouse in North America. *Condor*. 106(2):363–376.
- Shaftel R, Rinella DJ, Kwon E, Brown SC, Gates HR, Kendall S, Lank DB, Liebezeit JR, Payer DC, Rausch J, et al. 2021. Predictors of invertebrate biomass and rate of advancement of invertebrate phenology across eight sites in the North American Arctic. *Polar Biol*. 44(2):237–257.
- Smith IT, Knetter SJ, Svancara LK, Karl JW, Johnson TR, Rachlow JL. 2021. Overlap between sagebrush habitat specialists differs among seasons: implications for umbrella species conservation. *Rangeland Ecol Manage*. 78:142–154.
- Smith JT, Evans JS, Martin BH, Baruch-Mordo S, Kiesecker JM, Naugle DE. 2016. Reducing cultivation risk for at-risk species: predicting outcomes of conservation easements for sage-grouse. *Biol Conserv*. 201:10–19.
- Smith JT, Tack JD, Berkeley LI, Szczypinski M, Naugle DE. 2018a. Effects of rotational grazing management on nesting greater sage-grouse. *J Wildl Manage*. 82(1):103–112.
- Smith JT, Tack JD, Berkeley LI, Szczypinski M, Naugle DE. 2018b. Effects of livestock grazing on nesting sage-grouse in central Montana. *J Wildl Manage*. 82(7):1503–1515.
- Telfer MG, Hassall M. 1999. Ecotypic differentiation in the grasshopper *Chorthippus brunneus*: life history varies in relation to climate. *Oecologia*. 121(2):245–254.
- Thornton MM, Shrestha R, Wei Y, Thornton PE, Kao S, Wilson BE. 2020. Daymet: daily surface weather data on a 1-km grid for North America, Version 4. https://daac.ornl.gov/cgi-bin/dsvviewer.pl?ds_id=1840.
- Timmer JM, Aldridge CL, Fernández-Giménez ME. 2019. Managing for multiple species: greater sage-grouse and sagebrush songbirds. *J Wildl Manage*. 83(5):1043–1056.
- Traba J, Gómez-Catasús J, Barrero A, Bustillo-de la Rosa D, Zurdo J, Hervás I, Pérez-Granados C, García de la Morena EL, Santamaría A, Reverter M. 2022. Comparative assessment of satellite- and drone-based vegetation indices to predict arthropod biomass in shrub-steppes. *Ecol Appl*:e2707.

- U.S. Fish and Wildlife Service. 2015. 50 CFR Part 17 Endangered and threatened wildlife and plants; 12-month findings for petitions to list the greater sage-grouse (*Centrocercus urophasianus*) as threatened or endangered (No. Docket No. FWS-R6-ES-2015-0146). Washington, D.C.: U.S. Department of Interior. <https://www.govinfo.gov/content/pkg/FR-2015-10-02/pdf/2015-24292.pdf>.
- USGS (US Geological Survey). 2012. National Elevation Dataset (NED). <https://www.sciencebase.gov/catalog/item/4fcf8fd4e4b0c7fe80e81504>.
- Veblen KE, Nehring KC, McGlone CM, Ritchie ME. 2015. Contrasting effects of different mammalian herbivores on sagebrush plant communities. *PLoS One*. 10(2):e0118016–e0118016.
- Walker BL, Naugle DE, Doherty KE. 2007. Greater sage-grouse population response to energy development and habitat loss. *J Wildl Manage*. 71(8):2644–2654.
- Walston LJ, Cantwell BL, Krummel JR. 2009. Quantifying spatiotemporal changes in a sagebrush ecosystem in relation to energy development. *Ecography*. 32(6):943–952.
- Whelan CJ, Wenny DG, Marquis RJ. 2008. Ecosystem services provided by birds. *Ann N Y Acad Sci*. 1134:25–60.
- Wu T, Hao S, Kang L. 2021. Effects of soil temperature and moisture on the development and survival of grasshopper eggs in inner Mongolian grasslands. *Frontiers in Ecology and Evolution*. 9.